http://www.stockton-press.co.uk/jim

5-Chloropicolinic acid is produced by specific degradation of 4chlorobenzoic acid by *Sphingomonas paucimobilis* BPSI-3

AD Davison^{1,2}, DR Jardine³ and P Karuso³

Departments of ¹Biological Sciences; ³Chemistry, Macquarie University, Sydney, NSW 2109, Australia

We have previously shown that the bacterium *Sphingomonas paucimobilis* BPSI-3, isolated from PCB-contaminated soil, can degrade halogenated biphenyls, naphthalenes, catechols and benzoic acids. However, before such an organism can be used in bioremediation, it is important to characterise the degradation products and determine the degradation pathways to ensure that compounds more toxic or mobile than the original contaminants are not produced. In the degradation of 4-chlorobiphenyl, *S. paucimobilis* BPSI-3 produces a novel chlorinated picolinic acid. In this paper, we show that 4-chlorobenzoate is an intermediate in this degradation and, through ¹⁵N-labelling, that 5-chloropicolinate is the only nitrogenous metabolite isolated under the extraction conditions used. The position of the chlorine indicates that degradation of 4-chlorocatechol occurs exclusively *via* a 2,3-extradiol cleavage. These data allow us to postulate a more definitive catabolic pathway for the biodegradation of 4-chlorobiphenyl to 5-chloro-2-hydroxymuconic acid semialdehyde *via* 4-chlorobenzoate in *S. paucimobilis* BPSI-3.

Keywords: biphenyl degradation; biodegradation; bioremediation; chloropicolinic acid

Introduction

Polychlorinated biphenyls (PCBs) were first discovered to be ubiquitous environmental pollutants in the 1960s [3]. Subsequently, their manufacture has been terminated, although they still persist in many environments. Concern is growing over their toxic effects, specifically on marine mammals [31], as well as their role as possible oestrogen mimics (see [23] for example).

There are various techniques available for destruction of PCBs and PCB-containing wastes [17]. However, these will probably not be cost-effective for treatment of low concentration wastes typically found in soils and sediments. Furthermore, some technologies, such as incineration, may exacerbate pollution through the production of dioxins, compounds considered more toxic than the parent compound. Interest has now turned to the use of microorganisms for degradation of PCBs and other organochlorines (see for example [12]).

PCBs were first shown to be amenable to biodegradation by Ahmed and Focht [1]. Many microbes have now been isolated that are capable of degrading chlorinated and nonchlorinated biphenyls, generally via the intermediate chlorobenzoic acid. Some organisms, however, can exacerbate the toxicity of chlorinated biphenyls through the production of arene oxides as intermediates in the degradation process [11,30]. Furthermore, deleterious compounds, such as the antibiotic protoanemonin, are produced from other chloroaromatic compounds and are thought to contribute to their ecotoxicity [8,9,24]. Therefore, any biodegradation approaches to the remediation of contaminated sites should take into account not only the disappearance of the parent substrate, but also the appearance of catabolites. Furthermore, many microorganisms can only degrade chlorinated biphenyls to their corresponding chlorobenzoates. The first report of 4-chlorobiphenyl mineralisation to CO_2 and water *via* 4-chlorobenzoate was by Shields *et al* [29]. They found that mineralisation was mediated by the plasmid pSS50, located in several species of *Alcaligenes* and *Acinetobacter*. Exhaustive analytical procedures are therefore required not only to monitor biodegradation efficacy, but also to characterise interesting microbial isolates catabolically for their ability to degrade a variety of pollutants.

Sphingomonas paucimobilis BPSI-3 was isolated from a PCB-contaminated site in Balmain near Sydney [15]. It has a broad substrate range, being able to attack many halogenated aromatic structures including biphenyls, naphthalenes, catechols and benzoates [13,15]. In the course of these studies, S. paucimobilis BPSI-3 was found to produce a chloropicolinic acid from 4-chlorobiphenyl [14]. Gas chromatography-mass spectrometry (GC-MS) was used to identify this compound through comparison of the fragmentation pattern against library standards. A pathway was postulated for its formation from 4-chlorobenzoic acid based on known catabolic pathways although, from the available data, it was not possible to assign the position of the chlorine on the picolinate ring. Such stereochemical assignments of substituted muconic acid semialdehydes have previously been based on GC, GC-MS and UV characteristics of the products, but these techniques are notoriously insensitive to the identification of positional isomers.

In this paper, we describe the use of ¹⁵N-labelling experiments and nuclear magnetic resonance (NMR) spectroscopy to elucidate which isomer of chloropicolinic acid is produced by *S. paucimobilis* BPSI-3 from 4-chlorobenzoic acid. These data have allowed us to propose a more detailed pathway for catabolism of 4-chlorobenzoic acid by *S. paucimobilis* BPSI-3 based on unambiguous assignment

Correspondence: P Karuso, School of Chemistry, Macquarie University, NSW 2109, Australia

²Present address: Australian Water Technologies, 51 Hermitage Rd, West Ryde, NSW 2114, Australia

Received 19 April 1999; accepted 23 July 1999

348

of the chloropicolinic acid structure and, by inference, the structure of the precursor, muconic acid semialdehyde.

Materials and methods

Bacterial culture

Cultures of S. paucimobilis (24-40 h; 25°C) were grown and harvested as previously described [14,15] in PCBMS. a minimal medium composed of: solution A: 2.25 g $(NH_4)_2SO_4$ in 890 ml distilled water; solution B (g L⁻¹): K₂HPO₄, 44.0; KH₂PO₄, 17.0; pH 7.0 ± 0.02; solution C $(g L^{-1}):$ $MgSO_4 \cdot 7H_2O_2$ 19.5; $MnSO_4 \cdot H_2O$, 5.0; FeSO₄·7H₂O, 1.0; CaSO₄·2H₂O, 0.3; conc H₂SO₄, 500 µl (to prevent precipitation). Solutions A and B were autoclaved at 121°C for 15 min and solution C was filter sterilised $(0.22 \mu m)$. A 100-ml volume of solution B, 10 ml of solution C and the 890 ml of solution A were mixed after sterilising to make 1 L. The cell suspension was washed twice in PCBMS and added to 50 ml PCBMS (in a 250-ml Erlenmeyer flask) containing dextrose (1 g L^{-1}) and 4-chlorobenzoate (100 mg L^{-1}) to give a final absorbance at 600 nm of 0.4. The culture was incubated in the dark with shaking (25°C; 180 rpm). After 24 h, the whole culture was acidified to pH 2-3 using orthophosphoric acid (Analar, BDH, Sydney, Australia) and centrifuged (ca $10\,000 \times g$, 30 min, 4°C) to remove cells and debris. The supernatant was extracted with two volumes of ethyl acetate (HPLC grade, BDH).

The extract was dried over anhydrous sodium sulfate and concentrated to approximately 500 μ l *in vacuo* using gentle heat (30°C). The sample volume was reduced further under a stream of nitrogen and methylated with diazomethane in ether. Diazomethane was generated from *N*-methyl-*N*-nitroso-*p*-toluenesulfonamide using the Diazald[®]kit according to the manufacturer's instructions (Aldrich Chemical Company, Milwaukee, WI, USA).

¹⁵N-labelling conditions

Cultures of *S. paucimobilis* BPSI-3 were produced as described previously [14,15], with some modifications. The labelling experiment was carried out using [¹⁵N]-ammonium sulfate (99atom%, ICN, Aurora, OH, USA)-amended mineral salts' solution (PCBMS) [15]. The ammonium sulfate solution was filter sterilized (0.22 μ m, to minimise loss of the ¹⁵N label) then made up according to original methods [15]. Culture and extraction conditions were then followed as described above.

Chemical analyses

GC-MS analysis of the sample was performed on an MD800 GC-MS system (Finnigan MassLab, Manchester, UK) coupled to a Carlo-Erba GC8000 GC. The GC column was a SGE BPX5 length 25 m, internal diameter 0.22 mm, with a 0.25 μ m film thickness. The injection (2 μ l) was performed in the splitless mode at an injector temperature of 250°C. The column temperature was held at 40°C for 2 min then increased at 5°C min⁻¹ to 280°C where it was held for 5 min. The carrier gas was helium at an inlet pressure of 70 kPa. Mass spectra were obtained in the EI mode (70 eV), the source temperature was scanned in 2 s. Possible structures

were identified by comparison of the mass spectra with those of a mass spectral database (National Institute of Standards and Technology, Gaithersburg, MD, USA), previous results [14] and by analysis of fragmentation patterns.

Nuclear magnetic resonance spectroscopy

The NMR sample was prepared from the uniformly ¹⁵N-labeled, methylated extract (~2 mg) by dissolving it in CDCl₃ (0.5 ml; 99.96 atom%, Aldrich, Milwaukee, WI, USA) and filtered into an NMR tube (Wilmald, Buena, NJ, USA, PP527). The sample was degassed and equilibrated under an atmosphere of nitrogen.

NMR data were acquired on a Bruker DMX600 (600 MHz) NMR spectrometer at 30°C and processed using xwinNMR (version 1.3; Bruker). All NMR experiments were run with quadrature detection with an ¹H spectral width of 6614 Hz and a recycle delay of 1 s. Chemical shifts were referenced to internal TMS (0 ppm). High power ¹H $\pi/2$ pulses were determined to be 7.25 μ s and low power (for MLEV-17 spin lock) at 32.4 μ s. MLEV sequences were flanked with trim pulses at the same low power of 2 ms duration. ¹⁵N high power $\pi/2$ pulse was 39 μ s and a low power pulse of 118 μ s was used for GARP decoupling. Gradient pulses were delivered along the *z*-axis with a ratio of 35:15:25 of full power for selection of ¹⁵N-¹H zero and double quantum coherence using a 1024-step Gaussian program.

Data for simple 1D experiments were acquired using a single 60° pulse with a recycle delay of 1 s. Sixty-four K real points were acquired and zero filled to 128 K and then Gaussian multiplied for resolution enhancement. ¹⁵N filtered spectra were acquired using a 1 D version of the longrange zero and double quantum correlation experiment of Bax [7] but modified with gradient pulses [32] for better double quantum selection. Selection for ${}^{2}J_{\rm NH}$ coupling was achieved using a delay of 45 ms (J 11 Hz; 32 scans) and for ${}^{3}J_{\rm NH}$ using a delay of 500 ms (J 1 Hz; 2048 scans). No decoupling was used in t2 due to the antiphase nature of the signals at the end of the pulse sequence. FIDs (8 K data points) were processed for resolution enhancement using a strong Gaussian apodisation and zero filled to 32 K. The spectra were magnitude calculated after Fourier transformation. The same experiment (J 11 Hz; 1024 scans) was modified with a trailing TOCSY transfer [21] of 72 ms. GARP decoupling was used during t2 and the final spectrum was magnitude calculated as above.

The size of the ${}^{15}N{}^{-1}H$ coupling constants was measured by running the same experiment with and without decoupling during t2. When decoupling was used, an extra 1/2Jdelay [6] was incorporated after the last ${}^{15}N$ pulse to allow the antiphase signals to refocus. In this case, the final spectrum was not magnitude calculated but otherwise processed as above.

Proton and ¹⁵N assignments were achieved using the simple proton detected, zero and double quantum coherence method of Bax [6]. No decoupling was used during the acquisition and the original 2 K data points was zero filled to 4 K in f2. One hundred and twenty-eight experiments (four scans each) in t1 (1000 ppm) were zero filled to 512 points and processed using quadrature filtering.

Results and discussion

We have previously shown [14] that S. paucimobilis can degrade 4-chlorobiphenyl to produce chlorohydroxymuconic acid semialdehyde, 4-chlorophenol and many other metabolites, one of which was consistent with a chloropicolinic acid. The picolinic acid was not formed in sterile controls or if nitrogen was omitted from the medium. We repeated the previous experiments with 4-chlorobenzoic acid and extracting the metabolites with acidified ethyl acetate yielded a pale oil, which was methylated (CH₂N₂/ether) to facilitate analysis by GC-MS (Figure 1). 4-Chlorophenol (1) was the major metabolite, along with chlorohydroxymuconic acid semialdehyde (3) {methyl ester} and chloropicolinic acid (2) {methyl ester} confirming that 4-chlorobenzoic acid is an intermediate in the biodegradation of 4-chlorobiphenyl by S. paucimobilis BPSI-3. This conclusion is based on the fact that S paucimobilis BPSI-3 can degrade 4-chlorobenzoic acid and that we have previously isolated the same compounds starting from 4-chlorobiphenyl [14]. Two peaks were seen for chloro-2-methoxymuconic acid semialdehyde (3a, 3b) {methyl ester} which arise (presumably) from *cis-/trans*isomers about the 2,3-double bond, trapped by methylation of the enol form. In contrast, only one peak was seen for the non-methylated metabolite (3) {methyl ester} indicating that enolisation occurs rapidly, but slowly enough to cause broadening of the GC peak. The observed products are generally dead-end metabolites but the production of 4-chlorophenol (1) was indicative of initial 1,2-dioxygenation of 4chlorobenzoic acid, the product of which could readily lose CO_2 and water to yield 1 (Figure 2).

The peak eluting at 22.8 min (Figure 1) had a mass spectral fragmentation pattern (Figure 3a) very similar to 6-chloropicolinic acid methyl ester (National Institute of Standards and Technology). A molecular ion at m/z 171/173 (3:1) suggested the presence of one chlorine and one nitrogen (odd molecular weight). The peaks at m/z



Figure 1 Total ion chromatogram of the ethyl acetate extract from *S. paucimobilis* incubated for 24 h with [¹⁵N]-PCBMS medium and 4-chlorobenzoate. Prominent metabolites are 4-chlorophenol (1), chloropicolinic acid (2) methyl ester, 5-chloro-2-hydroxymuconic acid semialdehyde (3) methyl ester and *cis/trans*-5-chloro-2-methoxymuconic acid semialdehyde methyl ester (3a,3b). Peaks labelled with an * correspond to column bleed.

140/142 (M⁺⁻ – OCH₃) and m/z 112/114 (M⁺⁻–COOCH₃) confirmed the presence of a carbomethoxy group. Loss of HCl from the peak at 112 yielded a peak at 76 Da, consistent with a pyridine nucleus. The position of the carbomethoxy group was deduced from biogenic considerations (Figure 2) and confirmed by peaks in the mass spectrum at m/z 141/143 and 115/113 resulting from loss of OCH₃ and COOCH₃ respectively with concomitant proton transfer to the nitrogen. This can occur only when the carbomethoxy is in the 2-position. It was not possible to determine the position of the chlorine from mass spectral data.

Determining the position of the chlorine substituent is not trivial, but it can be done by NMR spectroscopy. In particular, heteronuclear (¹³C-¹H) long range coupling constants (HMBC) can be used. However, in this case, there were a number of factors precluding this approach. Most importantly, we wanted to identify all nitrogenous metabolites of 4-chlorobenzoate. This necessitated working with the crude extract, as any purification might remove minor metabolites or by-products. The background signals from the crude extract (eg see Figure 4a) would make interpretation of any HMBC experiment impossible. In addition, the tedious purification of the metabolities by either GC or HPLC would yield small amounts (<1 mg) of metabolites which would make the HMBC experiment difficult as its sensitivity is very low. To solve these problems, we introduced an ¹⁵N label which, like ¹³C, can be used to differentiate the isomers through chemical shift and coupling patterns, but is free from background signals.

Incubations of *S. paucimobilis* in ¹⁵N PCBMS with 4chlorobenzoate were essentially identical with those in normal PCBMS with the exception of the fragmentation pattern of the peak eluting at 22.8 min (Figure 3b). The shifting of all peaks to a heavier mass by 1 Da confirmed the presence of one N atom in the methyl ester of **2** (Figure 1) and the near-quantitative incorporation of ¹⁵N. Zero and double quantum NMR experiments on the crude extract were used to establish the position of the chlorine.

According to Kemp [19], the ¹⁵N chemical shift in pyridines is around 320 ppm displaying ${}^{2}J_{\rm NH}$ values around 12 Hz, ${}^{3}J_{\rm NH}$ around 1.5 Hz and ${}^{4}J_{\rm NH}$ values of 0.2 Hz [22]. In addition, we would expect ${}^{3}J_{\rm HH}$ couplings of around 8 Hz, ${}^{4}J_{\rm HH}$ of around 2.5 Hz and ${}^{5}J_{\rm NH} < 1$ Hz. This coupling pattern should make it possible to elucidate the structure of the picolinate, irrespective of the substitution pattern. We first used a 2 D experiment to correlate ¹⁵N to ¹H through heteronuclear long-range coupling. This revealed one ¹⁵N signal in the expected ¹⁵N chemical shift range, which was coupled to two proton signals at δ 8.70 and 8.10. No other signals were observed, suggesting that there was only one constitutional isomer of methyl chloropicolinate in our mixture. In addition, there were two weak signals at δ 6.69 and 6.52, coupled to one ¹⁵N signal at 123 ppm. The ¹⁵N and ¹H chemical shifts are typical of amides. A 1 D version of the 2 D experiment (Figure 4b) optimised for a coupling constant of 12 Hz revealed only one signal $(\delta_{\rm H} 8.70)$. The combination of chemical shifts $(\delta_{\rm H}/\delta_{\rm N})$ and coupling constant (11.0 Hz) identified the proton as H6 of a pyridine ring. Repetition of this experiment with inclusion of a refocussing delay after the last ¹⁵N pulse and decoupling (GARP) of ¹⁵N during the acquisition period (data not



Figure 2 Catabolic pathway for the degradation of 4-chlorobenzoate to 5-chloropicolinate. Decarboxylation and *meta*-cleavage are facilitated by dioxygenases $[O_2]$. Three possible modes of aromatic ring cleavage are possible: proximal extradiol (a); distal extradiol (b); or intradiol cleavage. Only pathway 'a', which involves cleavage of the 2–3 bond of 4-chlorocatechol, is found in *S. paucimobilis* BPSI-3. Compounds indicated in square brackets were not isolated but are postulated intermediates. The structure of (2) indicates the coupling constants measured from Figure 4a and the numbering system used in the text. Structures in square brackets were not identified but inferred as intermediates on biogenic grounds.

shown) collapsed the ¹H signal into a doublet ($J_{\rm HH}$ 2.4 Hz). Thus the larger coupling (11.0 Hz) observed in the ¹H NMR spectrum (Figure 4a) is due to the adjacent ¹⁵N and this proton (H6) has no adjacent proton. A zero/double quantum 1 D experiment optimised for a $J_{\rm HN}$ coupling of 1 Hz (Figure 4c) revealed a new signal at δ 8.10, a doublet $(J_{\rm HH} 8.3 \, \text{Hz}; \text{Figure 4a})$. The third proton in the spin system was revealed by appending a TOCSY period after the last ¹⁵N pulse of the refocussed zero and double quantum correlation experiment (Figure 4d), This spectrum contained the last two signals plus another at δ 7.84, which apparently had no coupling to ¹⁵N but was coupled to the protons resonating at 8.10 ppm ($J_{\rm HH}$ 8.4 Hz) and 8.70 ppm $(J_{\rm HH} 2.4 \text{ Hz})$. This pattern is consistent only with 5-chloropicolinic acid (Figure 2) where the signals at 8.10 and 7.84 ppm are ortho to each other ($J_{\rm HH}$ 8.3 Hz) and H6, adjacent to the ¹⁵N ($J_{\rm HN}$ 11.0 Hz) is separated from H4 (δ 7.84) by the chlorine at C5.

Nitrogen-containing heterocyclic compounds have previously been observed in the degradation of biphenyls. For example, Ahmed [2], postulated the presence of a 4-chlorophenyl-substituted picolinic acid as a degradation product of 4-chlorobiphenyl by some strains of Pseudomonas putida carrying cloned genes from P. testosteroni B-356. However, the picolinate still contained the chlorinated ring, and the substitution pattern was not elucidated. 2-Picolinic acid has also been reported as a metabolite of catechol [10] degradation and through the action of metapyrocatechase on 4-substituted catechols [25] and aminophenols [20]. In our case, the results are significant because the presence of 5-chloro-2-picolinate, aside from being a novel metabolite, demonstrates that S. paucimobilis BPSI-3 is capable of sequentially attacking the aromatic rings of 4-chlorobiphenyl—even the halogenated ring can, at least, be cleaved. In addition, the presence of the chlorine has, through ¹⁵N labelling, allowed us to define unambiguously the cleavage

350



Figure 3 Electron impact (70 eV) mass spectral fragmentation pattern of unlabelled (a) and [¹⁵N]-chloropicolinic acid methyl ester (b) obtained through GCMS of the methylated extract. All the isomers of chloropicolinate have very similar fragmentation patterns making them impossible to differentiate by mass spectrometry alone.

pathway from benzoic acid to 2-hydroxymuconic acid semialdehyde. Previous authors [28] have inferred this pathway but assignment of muconic acid isomers through GC, GC-MS or UV spectroscopy can be unreliable as isomers that differ only in the position of the halogen are not easily differentiated.

Picolinic acid production has been attributed to the abiotic reaction of 2-hydroxymuconic acid semialdehydes with ammonia [5,27]. However, E. coli HB107/pSLMK1, cloned with Pseudomonas catechol 2,3-dioxygenase (metapyrocatechase), did not produce picolinic acids under culture conditions but these were formed when concentrated ammonia (ca 5 M) was added and allowed to react for 6 days [5]. In contrast, we observed relatively high yields of picolinic acids under physiological conditions with no added ammonia outside that used in the PCBMS



¹H NMR spectra of the ¹⁵N-labelled extract showing the nor-Figure 4 mal proton NMR spectrum (a) of the crude extract, the zero and double quantum filtered spectrum optimised for a coupling of 11 Hz (b) cleanly reveals only H6 of the chloropicolinate (methyl ester). Trace (c) is optimised for a coupling of 1 Hz and shows H6 and H3. Finally, trace (d) is the same as (b) but with a trailling TOCSY transfer, showing all three protons on the pyridine ring.

medium, leaving open the possibility of enzyme-assisted formation of picolinates in S. paucimobilis BPSI-3. In other cases [4,20], the nitrogen is already present in the degradation of amino or nitro benzenes, which produce 2-aminomuconic acid semialdehydes and there is little doubt that the cyclisation and dehydration are spontaneous.

The environmental impact of releasing picolinic acids is unknown but deserves further research before organisms capable of PCB degradation are released, particularly as picolinic acids are known as dopamine β -hydroxylase inhibitors [26]. It is also important to note that although S. paucimobilis BPSI-3 may not be an efficient degrader of chlorobiphenyl in terms of mineralisation, it may prove to be a useful member of a bacterial consortium [16] that can achieve complete degradation of this and other organochlorines.

The data presented here show that 5-chloropicolinate is the only alkaloid produced by S. paucimobilis BPSI-3 from the degradation of 4-chorobiphenyl via 4-chlorobenzoic acid that can be extracted with ethyl acetate under acidic conditions. This degradation entails exclusive cleavage of the 2,3-bond in 4-chlorocatechol by the putative dioxygenase (pathway 'a', Figure 2). The resulting lactone (Criegee rearrangement [27]) would be easily hydrolysed to 5-chloro-2-hydroxymuconic acid semialdehyde, which was also observed (Figure 1), consistent only with a proximal extradiol cleavage pathway.

Acknowledgements

This work was partially funded by a NSW Environmental Trusts' Research Grant. ADD was funded by a Macquarie University Research Fellowship which is gratefully

351

acknowledged. Support from the Australian Research Council (to PK) enabled the purchase of the 600 MHz NMR spectrometer used in this study.

Note added in proof

Since submission of this manuscript, Katsivela *et al* [18] reported the production of 5-chloropicolinic acid from 4-chlorocatechol through a similar proximal extradiol cleavage pathway by a new bacterium (LW1). The authors used our [14] mass spectral arguments to propose a structure for the chloropicolinate and confirmed this using a $^{1}H^{-13}C$ HMBC NMR experiment. This approach has the disadvantages that the metabolite must be purified before analysis and that relatively large quantities are required to obtain long-range $^{1}H^{-13}C$ correlations at natural abundance.

References

- Ahmad M and DD Focht. 1973. Degradation of polychlorinated biphenyls by two species of *Achromobacter*. Can J Microbiol 19: 47–52.
- 2 Ahmed D, M Sylvestre, M Sondossi and R Massé. 1991. Bioconversion of 2-hydroxy-6-oxo-6-(4'-chlorophenyl)hexa-2,4-dienoic acid, the meta-cleavage product of 4-chlorobiphenyl. J Gen Microbiol 137: 1375–1385.
- 3 Anon. 1966. Report of a new chemical hazard. New Sci 32: 612.
- 4 Aoki K, S Takenaka, S Murakami and R Shinke. 1997. Partial purification and characterization of a bacterial dioxygenase that catalyses the ring fission of 2-aminophenol. Microbiol Rev 152: 33–38.
- 5 Asano Y, Y Yamamoto and H Yamada. 1994. Catechol 2,3-dioxygenase-catalyzed synthesis of picolinic acids from catechols. Biosci Biotechnol Biochem 58: 2954–2956.
- 6 Bax A, RH Griffey and BL Hawkins. 1983. Correlation of proton and nitrogen-15 chemical shifts by multiple quantum NMR. J Magn Reson 55: 301–315.
- 7 Bax A and MF Summers. 1986. Proton and carbon-13 assignments from sensitivity-enhanced detection of heteronuclear multiple-bond connectivity by 2 D multiple quantum NMR. J Am Chem Soc 108: 2093–2094.
- 8 Blasco R, RM Wittich, M Mallavarapu, KN Timmis and DH Pieper. 1995. From xenobiotic to antibiotic, formation of protoanemonin from 4-chlorocatechol by enzymes of the 3-oxoadipate pathway. J Biol Chem 270: 29229–29235.
- 9 Brückmann M, R Blasco, KN Timmis and DH Pieper. 1998. Detoxification of protoanemonin by dienelactone hydrolase. J Bacteriol 180: 400–402.
- 10 Dagley S, WC Evans and DW Ribbons. 1960. New pathways in the oxidative metabolism of aromatic compounds by microorganisms. Nature (London) 188: 560–566.
- 11 Daly JW, DM Jerina and B Witkop. 1972. Arene oxides and the NIH shift: the metabolism, toxicity and carcinogenicity of aromatic compounds. Experientia 28: 1129–1264.
- 12 Davison AD. 1998. Phenol colada: microbes party on toxic cocktails. Microbiol Austral 19: 24–26.
- 13 Davison AD, H Csellner, P Karuso and DA Veal. 1994. Synergistic

growth between two members in a mixed microbial consortium growing on biphenyl. FEMS Microbiol Ecol 14: 133–146.

- 14 Davison AD, P Karuso, DR Jardine and DA Veal. 1996. Halopicolinic acids, novel products arising through the degradation of chloro- and bromo-biphenyl by *Sphingomonas paucimobilis* BPSI-3. Can J Microbiol 42: 66–71.
- 15 Davison AD and DA Veal. 1993. Storage of a mixed microbial consortium capable of growth on biphenyl. Lett Appl Microbiol 17: 101–103.
- 16 Davison AD and DA Veal. 1997. Synergistic mineralisation of biphenyl by Alcaligenes faecalis type II BPSI-2 and Sphingomonas paucimobilis BPSI-3. Lett Appl Microbiol 25: 58–62.
- 17 Independent PoIW. 1992. A cleaner Australia. In: Assessment of the Management Options. Independent Panel on Intractable Waste (Australia), vol 2, Paddington NSW, Australia.
- 18 Katsivela E, V Wray, DH Pieper and R-M Wittich. 1999. Initial reactions in the biodegradation of 1-chloro-4-nitrobenzene by a newly isolated bacterium, strain LW1. Appl Environ Microbiol 65: 1405–1412.
- 19 Kemp W. 1986. NMR in Chemistry: A Multinuclear Introduction. McMillan, London. 188 pp.
- 20 Lendenmann U and JC Spain. 1996. 2-Aminophenol 1,6-dioxygenase: a novel aromatic ring cleavage enzyme purified from *Pseudomonas pseudoalcaligenes* JS45. J Bacteriol 178: 6227–6232.
- 21 Lerner L and A Bax. 1986. Sensitivity-enhanced two-dimensional heteronuclear relayed coherence transfer NMR spectroscopy. J Magn Reson 69: 375–380.
- 22 Levy GC and RL Lichter. 1979. Nitrogen-15 Nuclear Magnetic Resonance Spectroscopy. John Wiley and Sons, New York, pp 115–118.
- 23 Matthiesen P. 1996. Endocrine disruption in wildlife. International Symposium on Environmental Chemistry and Toxicology (INTERSECT 96), Sydney Australia.
- 24 Megharaj M, R Blasco, RM Wittich, DH Dietmar and KN Timmis. 1996. Ecotoxicity of some chloroaromatic pollutants due to microbial formation of protoanemonin. Microbiol Aust 17: A58.
- 25 Nozaki M, S Kotani, K Ono and S Senoh. 1970. Metapyrocatechase III. Substrate specificity and mode of ring fission. Biochim Biophys Acta 220: 213–223.
- 26 Piesche L, H Hilse, J Oehlke, E Schroetter and P Oehme. 1983. Substituted picolinic acids as DBH [dopamine β-hydroxylase] inhibitors. Inhibition of dopamine β-hydroxylase and antihypertensive effect. Pharmazie 38: 335–338.
- 27 Sanvoisin J, GJ Langley and TDH Bugg. 1995. Mechanism of extradiol catechol dioxygenase: evidence for a lactone intermediate in the 2,3-dihydroxyphenylpropionate 1,2-dioxygenase reaction. J Am Chem Soc 117: 7836–7837.
- 28 Seo DI, JC Chae, KP Kim, Y Kim, KS Lee and CK Kim. 1998. A pathway for 4-chlorobenzoate degradation by *Pseudomonas* sp S-47. J Microbiol Biotechnol 8: 96–100.
- 29 Shields MS, SW Hooper and GS Sayler. 1985. Plasmid-mediated mineralisation of 4-chlorobiphenyl. J Bacteriol 163: 882–889.
- 30 Sylvestre M, R Massé, F Messier, J Fauteux, JB Bisaillon and R Beaudet. 1982. Bacterial nitration of 4-chlorobiphenyl. Appl Environ Microbiol 44: 871–877.
- 31 Tanabe S. 1996. Global contamination by persistent organochlorines and their ecotoxicological impact on marine mammals. International Symposium on Environmental Chemistry and Toxicology (INTERSECT 96), Sydney Australia.
- 32 Willker W, D Leibfritz, R Kerssebaum and W Bermel. 1993. Gradient selection in inverse heteronuclear correlation spectroscopy. Magn Reson Chem 31: 287–292.

352